



**Canada's Wild Salmon Policy: an assessment of conservation progress in British Columbia**

Journal:	<i>Canadian Journal of Fisheries and Aquatic Sciences</i>
Manuscript ID	cjfas-2017-0127.R1
Manuscript Type:	Rapid Communication
Date Submitted by the Author:	27-Jun-2017
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Is the invited manuscript for consideration in a Special Issue? :	N/A
Keyword:	fisheries management, conservation status, overexploitation, STOCK ASSESSMENT < General, population recovery

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1 Canada's Wild Salmon Policy: an assessment of conservation progress in British Columbia  
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- 23 **Keywords:** fisheries management, conservation status, overexploitation, stock assessment,  
24 population recovery

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25 **Abstract**

26 Canada's Policy for Conservation of Wild Pacific Salmon has been heralded as a  
27 transformative approach to the management of wild salmon whereby conservation is the highest  
28 priority. Given that changes to the Policy are under consideration, it is timely that we understand  
29 whether our state of knowledge and the status of wild salmon in Canada have indeed improved  
30 after its adoption in 2005. To answer these questions, we used two indices of improvement: 1)  
31 monitoring effort, and 2) abundance of spawning adults. Our results, based on data for all species  
32 from British Columbia's north and central coasts, show that monitoring effort has continued to  
33 erode, abundance of spawning adults has significantly declined for several species, the status of  
34 many salmon Conservation Units are in zones of concern, and 42% of the Conservation Units  
35 that we assessed as Red (threatened) would have improved in status had the Canadian fishery  
36 been reduced. We conclude with recommendations to help improve our knowledge of the status  
37 of salmon, and enable a robust and successfully implemented Wild Salmon Policy for the future.

## 38 **Introduction**

39 Canada's Policy for Conservation of Wild Pacific Salmon, referred to as the Wild Salmon  
40 Policy (WSP; DFO 2005), provides an integrated approach to the management of wild salmon  
41 (*Oncorhynchus* spp.) in British Columbia (BC) and Yukon Territory. In brief, the WSP outlines  
42 the specific steps by which Canada's commitment to the precautionary principle is to be applied  
43 to the conservation of wild Pacific salmon (Cohen 2012). Salmon diversity is to be managed and  
44 protected at the level of the Conservation Unit (CU); these are genetically and/or geographically  
45 distinct populations which, if extirpated, are unlikely to recolonize naturally within a human life-  
46 time (DFO 2005). When introduced to Canadian society in 2005, the WSP was considered  
47 transformative and timely: transformative by setting out a new conservation ethic that placed its  
48 highest priority on the conservation of salmon above all other uses (Cohen 2012), and timely in  
49 that it was a much-needed policy developed in response to repeated criticism from key  
50 stakeholders and Canada's Auditor General (e.g., Office of the Auditor General 1999, 2004)  
51 based (in part) on the eroding abundance of salmon, salmon habitat, and information required to  
52 assess population health (Irvine 2009).

53 Population metrics quantified by Fisheries and Oceans Canada's (DFO) salmon stock  
54 assessment program are the fundamental building block of the WSP. The purpose of the stock  
55 assessment program is to provide relevant information on biological status, trends, and  
56 productivity required to guide the decision-making process related to salmon populations,  
57 fisheries, and conservation (English 2016). The backbone of such a program is the annual  
58 estimates of fish returning to spawn ("escapement"); the time series of which extends for more  
59 than 60 years for many salmon populations in BC. Some of the most obvious and compelling  
60 reasons for obtaining spawning escapement estimates are to: i) monitor the health of salmon

61 populations for conservation, ii) set and adjust fisheries management goals, iii) assess the impact  
62 of climate change, fisheries, and other human activities (e.g., logging, mining, etc.) on salmon,  
63 and iv) meet Canada's commitments to international treaties and First Nations. Despite such  
64 importance, monitoring effort for spawning streams had been in decline leading up to the  
65 adoption of the WSP in 2005, where 70% of all streams on BC's north and central coasts had not  
66 a single estimate of abundance (Price et al. 2008). While spawning streams are not the unit of  
67 conservation under the WSP, this dearth of information resulted in the inability to apply status  
68 evaluations to 41% of stream populations throughout BC (Slaney et al. 1996). The adoption of  
69 the WSP renewed optimism that monitoring effort for spawning streams would improve so as to  
70 provide relevant information on productivity (recruitments per spawner), trends in abundance,  
71 and biological status of wild salmon.

72 Wild salmon have been in a state of decline in BC for several decades. As of 1993, 600 of  
73 9,204 salmon runs were considered at high risk of extirpation, 63 at moderate risk, and 57 were  
74 of special concern; 105 stream populations throughout BC were documented as extirpated  
75 (Slaney et al. 1996). Ninety-six percent of monitored streams on BC's north and central coasts  
76 consistently failed to meet management escapement targets during 1950 to 2005 (Price et al.  
77 2008). In the Skeena watershed, Canada's second largest salmon producing system, roughly one  
78 third of the original biodiversity (as measured by the number of genetically distinct spawning  
79 units) is thought to have been lost before the 1950s due to habitat loss and heavy fisheries  
80 exploitation (Walters et al. 2008). Indeed, salmon in BC have been exploited for food for  
81 millennia, and by commercial industries since the late 19<sup>th</sup> century (Argue and Shepard 2005). A  
82 federal audit of DFO in 1999 reported that Pacific salmon fisheries were in trouble, stating that  
83 "The long-term sustainability of the fisheries was at risk because of overfishing, habitat loss, and

84 other factors” (Office of the Auditor General 1999). While commercial fishery catches between  
85 1995 and 2005 were at the time considered the lowest on record, catches since then (i.e., 2006 to  
86 2014) have further declined by nearly one-half (DFO 2017). For many CUs, this decline is the  
87 result of reduced salmon abundance, and increased conservation actions to protect these depleted  
88 stocks. For a few CUs, fisheries have not been permitted due to the reductions in monitoring  
89 efforts required to assess the status of stocks that once supported these fisheries.

90 Strategy 6 of the WSP commits to periodic performance reviews to determine what is,  
91 and what is not, working with the policy (DFO 2005). The last independent performance review  
92 occurred in 2011, and reported on the progress of implementing each action step. Of 17 action  
93 steps outlined in the WSP, 4 were rated as having been completed, and 13 were rated either  
94 partially completed or wholly incomplete (Gardner Pinfold 2011). Importantly, 2 of 3 action  
95 steps in Strategy 1 – standardized monitoring of wild salmon status – deemed critical to the  
96 overall success of the WSP (Cohen 2012), were reported as only partially complete. Six years  
97 have elapsed since the last performance review, and it is important to examine whether further  
98 progress has been made. Furthermore, changes to the WSP are being considered (DFO 2016a);  
99 thus, it is timely that we understand whether the policy in its current form ultimately has  
100 improved the health of wild salmon, their habitats, and dependent ecosystems in Canada.

101 The primary goal of our paper is to assess whether the state of our knowledge, and the  
102 biological status of wild salmon in Canada, has improved over the decade since the adoption of  
103 the WSP. To achieve this goal, we used 2 indices to assess improvement: 1) monitoring effort;  
104 whether monitoring effort of spawning streams had improved and whether a strategic approach  
105 to monitoring has been applied, and 2) spawner abundance; whether abundance of spawners has  
106 increased in CUs that were previously depressed, resulting in positive shifts in biological status

107 of CUs in BC. Three themes emerge from our provisional assessment: 1) the number of  
108 spawning streams assessed is at an all-time low, 2) there is inadequate information to determine  
109 the biological status of roughly one-half of all Conservation Units, and 3) implementation of the  
110 WSP needs to be given high priority. Given our results, we provide specific recommendations to  
111 improve our knowledge of salmon in BC, to ensure adequate protection is applied for diminished  
112 populations, and initiate a robust and successfully implemented WSP for the future.

### 113 **Methods**

114 We examined stream-specific escapement estimates between 1950 and 2014 (English  
115 2016), and run-reconstructed escapement and exploitation estimates for CUs between 1954 and  
116 2014 (English et al. 2016) for BC's north and central coasts (DFO Management Areas 1 through  
117 10; Fig. 1); similar estimates were not publicly available for BC's south coast salmon CUs. A  
118 complete list and the location of all CUs is reported in Holtby and Ciruna (2007). Briefly,  
119 escapement estimates for each CU were derived by expanding the available estimates for  
120 indicator streams – spawning streams considered biologically representative of the productivity  
121 across a given CU - within each CU. The expansion accounted for any missing estimates for  
122 indicator streams, the portion that the indicator streams represent of the total average escapement  
123 for the CU, and the tendency for estimates to underestimate escapements based on visual surveys  
124 (English et al. 2016). Exploitation rate estimates were derived using several different approaches  
125 depending on the species and CU, and are thoroughly described in English et al. (2016) and our  
126 Supplementary material. Despite the many assumptions underlying the run-reconstruction data  
127 (see Appendix E, English et al. 2016; our Supplementary material), the resulting uncertainty is  
128 unlikely to lead to systematic bias that will alter our ultimate inference because such uncertainty  
129 applies across the time series comparison detailed below. However, we acknowledge the

130 uncertainty associated with our assessment of how reduced fishing pressure could change the  
131 biological status of CUs between time periods. We assessed the five major salmon species in  
132 Canada: *O. tshawytscha* (Chinook), *O. keta* (chum), *O. kisutch* (coho), *O. gorbuscha* (pink), and  
133 *O. nerka* (sockeye). Given their distinct 2-year life cycle, we separated pink salmon into even-  
134 and odd-years for all analyses. There are 2,933 documented salmon spawning streams on BC's  
135 north and central coasts, many of which are small with less than a few hundred spawners of a  
136 given species, but may account for a disproportionate amount of the genetic diversity among  
137 populations (Hyatt et al. 2007).

### 138 **Monitoring effort**

139 Annual estimates of returns of each species to each management area and CU are derived  
140 from data collected during spawning escapement surveys, and stored in DFO's Salmon  
141 Escapement Database System (NuSEDS; DFO 2016b). While the accuracy of escapement  
142 estimates within this database system has been questioned (e.g., Irvine and Nelson 1995), there  
143 are few alternative data available. A sub-set of spawning streams consistently enumerated over  
144 time has further been classified as "indicator streams" based on historical time series, the  
145 reliability of escapement estimates, and the methods and costs of obtaining these data (English et  
146 al. 2006, 2016; Walters et al. 2008; Ogden et al. 2015). Our assessment of changes to monitoring  
147 efforts over time occurred at three scales: 1) total streams, 2) indicator streams, and 3) CUs. We  
148 used a linear regression to quantify the rate of change in monitoring effort of indicator streams  
149 for all species since the adoption of the WSP in 2005. For CUs, we examined the number of CUs  
150 with an assigned indicator stream, then examined monitoring effort by calculating the proportion  
151 of indicator streams surveyed for each of those CUs in all years during 2005-2014.

152 We also assessed whether fisheries managers used a strategic approach to the

153 enumeration of salmon in spawning streams. The World Summit on Salmon (WSS 2003)  
154 identified various concerns regarding Canada's stock assessment programs for Pacific salmon.  
155 These concerns led to the development of the Core Stock Assessment Program (CSAP), a DFO  
156 commitment to strategic monitoring of spawning populations for each species returning to BC's  
157 north and central coast (English et al. 2006). The CSAP identified a set of indicator streams for  
158 each stock group, and three primary monitoring activities were recommended: escapement,  
159 fishery, and productive capacity. We compared CSAP recommendations with recent monitoring  
160 efforts to determine the extent to which the strategic monitoring program was implemented (see  
161 English 2016).

## 162 **Population trends**

163 We calculated the difference in arithmetic average (geometric mean also calculated for  
164 comparison; Holt et al. 2009; our Supplementary material) spawner escapement by species for  
165 each CU in the decade *before* (1995-2004) and *after* (2005-2014) the adoption of the WSP to  
166 determine the percent change in spawner abundance. We tested whether the percent difference  
167 was significant using a Wilcoxon Signed-Rank Test for non-normal data (set the "paired"  
168 argument = TRUE). While we excluded, from all comparisons, CUs with <50% of years with  
169 escapement data within a decadal period, the number of years in a given decade at times differed  
170 for some CUs (e.g., a CU with 10/10 data-years in one decade may have had 7/10 data-years in  
171 the following decade). This occurred for 25 CUs; 11 during pre-WSP years, and 14 for post-  
172 WSP years. A sensitivity analysis for all CUs with missing years showed that differences in  
173 spawning abundance were never large enough to change the biological status of a CU (our  
174 Supplementary material).

175 The first strategy of the WSP states that the conservation status of salmon CUs must be

176 determined against specific biological benchmarks, such as spawner abundance, using a “stop-  
177 light” approach (i.e., ‘Green’, ‘Amber’, and ‘Red’ status zones; DFO 2005). While the WSP does  
178 not dictate any particular metric to assess the biological status of CUs, several examples are  
179 provided, and have subsequently been evaluated (e.g., Holt et al. 2009; Peacock and Holt 2010;  
180 Holt and Bradford 2011). However, to date, DFO has not published the biological status of CUs  
181 in BC. Our intention is not to perform this task for DFO, but rather to use a method that we  
182 consider reasonable for comparing the status of as many CUs and species across BC before and  
183 after the WSP was published. We have chosen to use the percentile approach for several reasons:  
184 i) they can be applied to a large number of data-limited CUs where reliable spawner escapement  
185 estimates are not available to derive stock-recruitment based benchmarks, ii) they have been  
186 applied to represent bounds (i.e., reference range) for management escapement goals (Pestal and  
187 Johnston 2015), and iii) an evaluation of percentile-based benchmarks in the context of other  
188 WSP abundance benchmarks currently is in progress (C. Holt pers. comm). We defined upper  
189 and lower benchmarks so as to delimit status zones by employing the 25<sup>th</sup> and 75<sup>th</sup> percentile of  
190 historic spawner abundance (1954-1994 for all species except Chinook, whose data spanned  
191 1985-1994), for each salmon CU. If a CU’s abundance in a given decade was below the 25<sup>th</sup>  
192 percentile, between the 25<sup>th</sup> and 75<sup>th</sup> percentiles, or above the 75<sup>th</sup> percentile, of the long-term  
193 abundance, the CU would be assigned “Red”, “Amber”, and “Green” status, respectively. The  
194 resulting status for each CU was compared between the decadal periods *before* (1995-2004) and  
195 *after* (2005-2014) the adoption of the WSP to determine a change in overall status.

196 We were interested to know whether or not resource managers have applied a more  
197 precautionary approach to fisheries management over the decade since the adoption of the WSP,  
198 and whether this approach resulted in changes to the status of CUs. We examined changes in

199 arithmetic average exploitation from Canadian fisheries for the periods *before* (1995-2004) and  
200 *after* (2005-2014) the adoption of the WSP, and used a Wilcoxon Signed-Rank Test for non-  
201 normal data (set the “paired” argument = TRUE) for each species to determine statistical  
202 significance. While some BC salmon stocks are caught in Alaskan fisheries, we excluded this  
203 exploitation from our analyses because it is beyond the control of resource managers in Canada.  
204 To examine the influence of exploitation on biological status, we performed three assessments:  
205 1) the change in Canadian fisheries exploitation between decadal periods for CUs assigned Red  
206 status in the decade post-WSP, 2) the number of CUs that would have declined in status had  
207 Canadian fisheries not been reduced post-WSP - we examined those CUs whose status did not  
208 change post-WSP, subtracted the average number of fish that would have been caught had  
209 exploitation not been reduced for a given CU, and compared the “revised” return to its  
210 benchmarks), and 3) the number of CUs that would have improved in status had Canadian  
211 fisheries been reduced from 0% to 100% on each CU - we assigned status for each CU post-WSP  
212 based on the total return to Canada estimates (Canadian catch plus escapement), compared the  
213 results with status derived from spawner abundance alone over the same period to determine  
214 whether status would have improved, and assessed the rate of exploitation that a given CU could  
215 sustain before its status declined. Two assumptions apply: 1) fish would not have experienced  
216 significant en-route mortality during upriver migration had they been allowed to escape fisheries  
217 capture, and 2) productivity is the same regardless of differences in numbers of spawning adults.

218 Finally, we performed a provisional status assessment for all CUs on BC’s north and  
219 central coasts for the contemporary period using the percentile approach stated above, and the  
220 arithmetic mean (geometric mean also calculated for comparison; Holt et al. 2009; our  
221 Supplementary material) of the following “generational” (i.e., over the most recent generation)

222 years: Chinook (2009-2014); chum (2011-2014); coho (2011-2014); even-year pink (2014); odd-  
223 year pink (2013); sockeye (2009-2014 depending on CU; English et al. 2016). Admittedly, there  
224 are uncertainties aligning percentile-based benchmarks with those defined under the WSP for  
225 data-rich CUs; the 25<sup>th</sup> and 75<sup>th</sup> percentile of abundance delineating Red and Green zones may be  
226 lower or higher than other metrics. We provide a cursory examination of differences in  
227 benchmark values (and associated status) between the percentile approach and a stock-  
228 recruitment approach in our Supplementary material. While we acknowledge that a status  
229 assessment integrated over numerous metrics is preferred, we believe that our provisional  
230 assessment based on a single metric serves to provide a rapid evaluation for resource managers  
231 of where conservation concerns may exist. All analyses and graphics were performed in R 3.3.2  
232 (R Foundation for Statistical Computing, 2016) using the cowplot, dplyr, ggplot2, gridExtra, and  
233 lm packages.

## 234 **Results**

### 235 **Monitoring effort**

236 The number of spawning streams with escapement estimates on BC's north and central  
237 coasts has varied widely, peaking in the mid-1980s at 1,533 streams, declining to less than 1,000  
238 by the mid-1990s, and reaching an all-time low of 476 streams in 2014 (Fig. 2a). Spawning  
239 locations referred to as "indicator streams" have been monitored more consistently over time.  
240 Escapement surveys averaged 490 (72%) indicator streams per year during 1950 to 2004, but has  
241 since declined at an average rate of 3.8% per year (for a total decline of 34% over the last 10  
242 years;  $R^2 = 0.614$ ,  $df = 7$ ,  $p = 0.013$ ) to an all-time low of 334 streams in 2014; thus, only 49% of  
243 the 679 indicator streams were surveyed in the most recent year (Fig. 2b). When monitoring  
244 effort was evaluated at the CU level, 58% (127 of 218) of CUs have at least one assigned

245 indicator stream (Table S1). Of those CUs with an assigned indicator stream, the median  
246 proportion of CU-specific stream visits during 2005-2014 ranged from 60% for Chinook to 80%  
247 for sockeye, and averaged 68% across all CUs over the decade since the WSP (Fig. 3).

248 Of the 634 annual, and 134 periodic, spawning streams recommended for monitoring by  
249 CSAP, only 29% were monitored consistently during the 2007-2014 period (Table 1). Coho  
250 streams were monitored the least (20%), whereas streams with lake-type sockeye were  
251 monitored the most (42%). Twenty-four percent of CSAP recommended streams had zero effort  
252 over the time period for all species combined, including a low of 56% for river-type sockeye.

### 253 **Population trends**

254 The percentage of CUs with spawner averages that had changed since the adoption of the  
255 WSP varied by species. The relative change was highest and statistically significant (Wilcoxon  
256 rank score,  $W = 146$ ,  $p = 0.000$ ) for chum salmon where post-WSP spawners were 23% of the  
257 average that returned during the decade prior to the WSP (Table 2; Fig. 4). Even-year pink ( $W =$   
258  $28$ ,  $p = 0.016$ ) and Chinook ( $W = 74$ ,  $p = 0.048$ ) also experienced large and statistically  
259 significant declines of 86% and 69%, respectively. Overall, there was a 14% decrease in average  
260 annual spawners in the decade after the WSP, driven largely by even-year pink salmon.

261 Of 218 CUs on BC's north and central coasts, 30 declined in biological status (e.g.,  
262 Green to Amber, or Amber to Red), 15 improved, and 15 changed to Unknown status since the  
263 adoption of the WSP (Fig. 5; Table S2). There was variation among species: Chinook and chum  
264 salmon had the highest proportion of CUs that declined in status to Red (25% and 24%,  
265 respectively). Alternatively, coho salmon had the highest proportion of CUs that improved in  
266 status to Green (21%), followed by odd-year pink salmon (17%). The total number of CUs for all  
267 species with Unknown status – most of which were small coastal sockeye CUs - increased from

268 100 to 108; thus, roughly one-half of all CUs on BC's north and central coasts had insufficient  
269 data to determine status in the post-WSP period.

270 The relative change in exploitation before and after the adoption of the WSP was highest  
271 (-59%) and statistically significant ( $W = 26$ ,  $p = 0.047$ ) for even-year pink salmon, followed by  
272 chum (-56%;  $W = 136$ ,  $p < 0.001$ ), and sockeye (-50%;  $W = 630$ ,  $p < 0.001$ ); Chinook  
273 experienced the least (-12%) change in exploitation (Fig. 6). The change in Canadian fisheries  
274 exploitation between decadal periods for those CUs assessed as Red in the decade post-WSP  
275 ranged from -11% (chum) to +1% (coho), and averaged -7% across species (Table 2). Regarding  
276 biological status, 3 CUs would have declined in status from Amber to Red had fishing pressure  
277 not been reduced over the decade since the adoption of the WSP. However, 10 CUs would have  
278 improved in status either to Amber or Green had Canadian fisheries exploitation been further  
279 reduced by 50%. Ten of 24 CUs that we assessed as Red would have improved in status; 4 of 6  
280 Chinook CUs would no longer be in the Red zone (Fig. 7).

281 Our contemporary period status assessment shows that only 5% of chum, 12% of  
282 Chinook, and 15% of sockeye CUs have Green status up to 2014; coho had the highest  
283 percentage of CUs with Green status (42%; Fig. 8; Table S2). Sixty-five percent of sockeye CUs  
284 are considered Unknown, and 50% of all CUs on BC's north and central coasts are of Unknown  
285 status in the contemporary period.

## 286 Discussion

287 Canada's Policy for Conservation of Wild Pacific Salmon has been articulated as the  
288 means by which the federal government will meet its obligation to protect and conserve wild  
289 salmon on the Pacific coast (Cohen 2012). We have asked whether the adoption of the WSP in  
290 2005 has improved our state of knowledge and the status of these iconic fish. While considerable

291 progress has been made in the 12 years since its adoption, implementation of the WSP is still far  
292 from complete. Three themes emerge from our assessment: 1) the number of spawning streams  
293 assessed is at an all-time low, 2) there is inadequate information to determine the biological  
294 status of roughly one-half of all Conservation Units, and 3) implementation of the WSP must be  
295 given high priority. To reverse these trends, and initiate a robust and successfully implemented  
296 WSP for the future, we conclude with specific recommendations.

297         Annual estimates of spawning salmon are the fundamental building block of fisheries  
298 management in Canada, essential for monitoring conservation status and estimating the total  
299 annual returns for each salmon Conservation Unit (CU; English et al. 2016). Such importance  
300 has long been acknowledged. For example, DFO's 1987 operational framework for Management  
301 Area 6 states that, "Escapement data are the basis of the whole fisheries management  
302 regime...neither pre-season planning nor computer modeling and run reconstruction or any other  
303 long-term strategic planning exercise is possible without this information." (DFO 1987). Despite  
304 its immense importance, visits to spawning streams on BC's north and central coasts have been  
305 trending downward since the mid-1980s; total stream visits in 2014 were 69% lower than those  
306 in 1986. Importantly, spawning locations referred to as indicator streams experienced a 34%  
307 reduction in effort since the adoption of the WSP in 2005. Indicator streams were selected by  
308 regional biologists because escapement estimates for these streams were more reliable and more  
309 consistently surveyed than those for other streams in a CU, and also because these streams were  
310 considered biologically representative of the productivity across a CU. Our state of knowledge  
311 regarding salmon populations is eroding rapidly. Without increased support for escapement  
312 surveys and the transfer of knowledge to the next generation, the rich legacy of population data  
313 available for BC's north and central coasts is at serious risk of becoming irrelevant for future

314 assessments of management and conservation status.

315           How does such monitoring effort translate to the CU level of fisheries management? We  
316 are unable to assess the status of 40% of all north coast and central coast salmon CUs because  
317 these CUs do not have an assigned indicator stream; many of these are small isolated sockeye  
318 CUs without other nearby sockeye CUs with indicator streams. In addition to these unmonitored  
319 CUs, there are significant gaps in the escapement data for CUs with indicator streams.  
320 Escapement estimates are not available for 32% of the indicator stream-years in the post-WSP  
321 decade. These deficiencies in escapement monitoring efforts leave fisheries managers with  
322 inadequate information to assess population health, and opportunities for local fisheries.

323           While budget shortfalls have contributed to monitoring declines, there also has been a  
324 lack of strategic approach towards monitoring that otherwise could have improved our  
325 knowledge state for data-deficient CUs. There are three inter-related short-comings. First, the  
326 implementation of the Core Stock Assessment Program - a strategic approach to annual  
327 escapement surveys - has fallen far short of its goals (English 2016). Only 29% of all indicator  
328 streams recommended to be monitored were surveyed consistently during the period 2007-2014,  
329 and 24% of the indicator streams had zero survey effort. Second, enumeration monitoring at the  
330 CU-level is highly variable, with several CUs having received exhaustive effort, while other CUs  
331 for the same species have been ignored completely (see Table S1). Finally, visits to indicator  
332 streams declined to their lowest level ever in 2014. On average, 150 non-indicator streams were  
333 enumerated annually since the WSP, when, had a strategic approach been followed, they need  
334 not have been. Had managers chosen to annually enumerate 150 more indicator streams, rather  
335 than non-indicator streams, monitoring effort for indicator streams would have been reinstated to  
336 the peak period levels of the 1980s, and far fewer CUs would now be considered data-deficient.

337 Our assessment reveals that salmon abundance has declined over the decade since the  
338 WSP, driven largely by even-year pink salmon. Climate variability and poor marine survival  
339 have played a substantial role in the diminishment of populations. For example, sockeye salmon  
340 throughout southern portions of their range have exhibited downward trends in productivity in  
341 recent decades (Peterman and Dorner 2012), resulting in part from competition with increasingly  
342 abundant pink salmon across the North Pacific (Ruggerone and Connors 2015). There also have  
343 been widespread declines in chum salmon throughout BC, and pink salmon more recently in  
344 several areas on BC's central coast, likely due to large-scale climatic processes (Malick and Cox  
345 2016). The overall decrease in average annual spawners that we report in the decade after the  
346 WSP was driven largely by the decline in even-year pink salmon, despite notable reductions in  
347 fisheries pressure. Similar to our results, recent increases in odd-year pink abundance in southern  
348 BC have been shown to be correlated with decreased fishery exploitation (Irvine et al. 2014),  
349 which begs the question as to what factor(s) may be influencing the differential dominance  
350 between odd- and even-year pink salmon. Evidence provided by Irvine et al. (2014) suggests that  
351 recent climate conditions may be challenging even-year pink salmon more than odd-year pink,  
352 due to their historical dispersal from divergent glacial refugia.

353 Resource managers have responded positively, at a broad scale, to diminished salmon  
354 returns over the last decade by reducing exploitation on all species in ocean fisheries, though not  
355 for some vulnerable CUs. If fishing pressure had not been reduced, 3 CUs (1 each of chum,  
356 even-year pink, and sockeye) would have declined in status from Amber to Red – assuming no  
357 significant change in productivity with these slightly higher numbers of spawning adults. More  
358 importantly, though, 10 of 24 CUs that we assessed as Red since the WSP would have improved  
359 in status to either Amber or Green had the Canadian fishery been further reduced, and all but 2

360 Chinook CUs would no longer be in the Red zone, assuming that all fish escaping the fishery  
361 were successful spawners; and these results are, of course, sensitive to assumptions in the run  
362 reconstruction data. A broad-scale reduction in Canadian fisheries exploitation from 40% to 30%  
363 would have improved the status of three CUs from Red to Amber or Green. However, fisheries  
364 exploitation would need to have been further reduced to <10% to improve the status of the  
365 majority of CUs in the Red zone, unless such exploitation were moved upriver to more terminal  
366 locations where vulnerable populations can be avoided. The degree to which exploitation rates in  
367 Canadian fisheries can be further reduced, or moved upriver, to improve the status for a few CUs  
368 is the subject of the type of trade-off discussion that the Skeena Independent Science Review  
369 Panel recommended nearly a decade ago (Walters et al. 2008).

370 Canada's Wild Salmon Policy provides the blueprint to safeguard the natural diversity of  
371 salmon, but slow progress towards defining the WSP benchmarks for salmon CUs has impeded  
372 the delivery of biological status assessments required to guide fisheries management. The  
373 classification of lower and upper benchmarks for exploited populations is an important action  
374 step for implementing the WSP, outlined in Strategy 1. While numerous candidate metrics have  
375 been proposed for data-rich CUs (e.g., Holt et al. 2009; Peacock and Holt 2010; Holt and  
376 Bradford 2011), benchmark development remains in the evaluation stage for some data-limited  
377 CUs (most of which occur along BC's north and central coasts). We understand that benchmark  
378 metrics will continue to evolve as new data are collected on CUs in BC, and that their evaluation  
379 is an ongoing objective of the WSP. However, Canada's management agency arguably has  
380 sufficient scientifically-defensible metrics to immediately assess the biological status of dozens  
381 of CUs throughout BC, especially those in the Skeena and Fraser watersheds. Furthermore,  
382 despite being a somewhat poor surrogate for stock-recruitment based benchmarks of data-rich

383 CUs (our Table S3), simple metrics such as the percentile approach can provide a rapid  
384 evaluation of conservation concerns for data-limited CUs, where stock-recruitment based  
385 benchmarks cannot be derived or are inappropriate. It is now 12 years since its publication and  
386 the WSP remains only partially implemented, with relatively few CUs having been formally  
387 assessed.

388         Can our contemporary period status assessment inform resource managers of  
389 conservation concerns? We believe so, in three ways. First, our results provide a “first cut” of  
390 where conservation concerns may exist. While such provisional assessments are not as rigorous  
391 as the recommended multi-metric approaches used for data-rich CUs (e.g., Fraser sockeye),  
392 assessments performed based on a single metric for dozens of CUs across a large region can  
393 identify where to prioritize more in-depth assessment efforts, including the need for a wider  
394 variety of data, metrics, and expertise. Second, patterns of diminished CUs are consistent across  
395 species. Our results show that, of those CUs with sufficient data to determine status, all species  
396 have 1 or more CUs assessed as Red and less than 50% assessed as Green. Third, and perhaps  
397 most important, one half of all CUs on BC’s north and central coasts have insufficient data to  
398 assign status using our proposed metric, and no alternative benchmark approach could markedly  
399 increase the number of CUs with assigned status. The take-home message to resource managers,  
400 is thus: conservation actions are required for each species, and more data need to be acquired to  
401 understand the true scale of the conservation concern for these iconic fishes. A logical next step  
402 would be to perform an analysis to determine which areas (or groups of CUs) host  
403 disproportionate declines in abundance or Red status, and whether exploitation has changed for  
404 those areas, so as to inform managers of where next to implement full integrated biological  
405 assessments.

406 To ensure adequate protection is applied for diminished populations, and to initiate a  
407 robust and successfully implemented WSP for the future, we make the following  
408 recommendations to the federal government of Canada and Canada's Department of Fisheries  
409 and Oceans:

410

411 **1. Conduct a strategic planning review of Conservation Units to meet the requirements of**  
412 **the Wild Salmon Policy.** Such a review should incorporate potential partners and collaborators  
413 to aid data acquisition and control financial cost, and restore key assessment programs as a  
414 priority for DFO annual programming. Implementation of the updated north and central coast  
415 core stock assessment program detailed in English (2016) would be an appropriate beginning to  
416 ensure that the most critical data for salmon management are collected each year, and provide the  
417 information necessary to adequately determine the biological status for most CUs on BC's north  
418 and central coasts. The total estimated annual cost for implementation is \$2.5 million, with an  
419 additional \$400,000 per year for 5 years recommended to build enumeration monitoring capacity  
420 - \$1.2 million more than funding allocated in recent years (English 2016). If funding is not  
421 improved for monitoring and status assessments of data-limited CUs, it is likely that a risk-based  
422 prioritization process will occur; in which case, we recommend that there be clear documentation  
423 of factors (e.g., conservation, First Nations and international obligations, habitat threats, etc.)  
424 considered in the prioritization.

425

426 **2. Use a two-step approach to speed up the process for assessing biological status.** Several  
427 candidate benchmark approaches have been identified for biological status assessments (Holt et  
428 al. 2009; Peacock and Holt 2010; Holt and Bradford 2011), and evaluated based on simulations

429 that have quantified extinction and recovery probabilities (Holt and Folkes 2015). Stock-  
430 recruitment, recent trend in spawner abundance, exploitation rate, rearing habitat capacity, and  
431 the percentile approach have all been proposed. Each has its limitation, and the accuracy of  
432 assessed status is derived from the integration of the metrics collectively. We recommend a two-  
433 step process towards status assessments: 1) use the percentile approach as an efficient initial  
434 region-wide assessment of stock status, and 2) where possible and appropriate, immediately  
435 integrate (see Grant and Pestal 2013) the information from a larger suite of metrics to increase  
436 the confidence in the assessments for CUs initially classified as Red and Amber. We caution,  
437 however, that this approach should not supersede the need for integrated status assessments of  
438 CUs initially classified as Green, but locally identified as of concern.

439  
440 **3. Achieve a balance between mixed-stock ocean fisheries and in-river fisheries targeting**  
441 **specific stocks.** This is a mitigation strategy that resource managers can use, and have used, in  
442 large watersheds like the Skeena and Fraser to address conservation concerns for specific CUs.  
443 A key requirement for implementing this strategy will be defining an initial set of management  
444 goals and benchmarks for the various CUs, and achieving some level of agreement between First  
445 Nations, recreational, and commercial fishing communities that catch fish returning to these  
446 watersheds. These management goals and benchmarks can be further refined over time with  
447 information from consistently applied escapement monitoring and other stock assessment  
448 programs.

449  
450 **4. Implement the existing WSP immediately.** The development of an implementation plan was  
451 assured in the WSP (DFO 2005), but has yet to be developed. Without further delay, a detailed

452 implementation plan should be developed that would stipulate the tasks required - how they will  
453 be performed, what collaborations and partnerships could assist in the tasks, and when they will  
454 be completed - and include a detailed breakdown of implementation costs, as Cohen (2012)  
455 recommended 5 years ago. While DFO recently has initiated a process to develop a plan, five  
456 years are proposed for implementation, together with changes to the WSP (DFO 2016a). We  
457 recommend that such a plan should immediately implement the WSP as written in 2005, thus  
458 retaining all Strategies and Action steps, as they are the blueprint for implementation that will  
459 ensure accountability. Indeed, “Further reviewing, reexamining, or reopening of the policy  
460 [WSP] would be a poor use of limited funds in the Pacific Region.” (Think Tank of Scientists  
461 2017).

462  
463 **5. Create a Wild Salmon Policy fund to ensure implementation.** While the WSP is a federal  
464 policy, the Pacific Region of DFO has thus far been responsible to find the funds within its own  
465 annual allocation to implement it (Cohen 2012). With the current federal government’s interest  
466 in First Nations and support for science, we recommend the creation of a WSP implementation  
467 fund to support the development of partnerships and the restoration of habitats where recovery of  
468 CUs is required. If DFO will not support annual needs for assessment and addressing CUs in the  
469 Red Zone as the WSP requires, the federal government could establish a fund managed  
470 collaboratively with DFO, yet directed via a trust fund.

471  
472 Canada’s Wild Salmon Policy sets out the specific steps by which Canada’s commitment  
473 to the precautionary principle is to be applied to the conservation of Pacific wild salmon. Our  
474 results show that monitoring effort has continued to erode, there is inadequate information to

475 determine the biological status of roughly one-half of all Conservation Units, and that an  
476 implementation plan for the Wild Salmon Policy is required now more than ever. Our five  
477 recommendations would help to ensure that Pacific salmon remain abundant in British Columbia  
478 for future generations.

#### 479 **Acknowledgements**

480 We thank the many biologists who have gathered and processed salmon data on BC's  
481 coast over the past 60 plus years, the Pacific Salmon Foundation and LGL Limited for compiling  
482 and reconstructing CU-specific data used in this manuscript, Sean Godwin for analytical advice,  
483 and Brendan Connors, Eric Hertz, Carrie Holt, Brian Riddell, and two anonymous reviewers for  
484 helpful comments. Funding for this project was provided by a Natural Sciences and Engineering  
485 Research Council of Canada (NSERC) Discovery Grant, and the Tom Buell Endowment Fund  
486 supported by the Pacific Salmon Foundation and the BC Leading Edge Endowment Fund (J.  
487 Reynolds), and an NSERC Alexander Graham Bell Canada Graduate Scholarship (M. Price).

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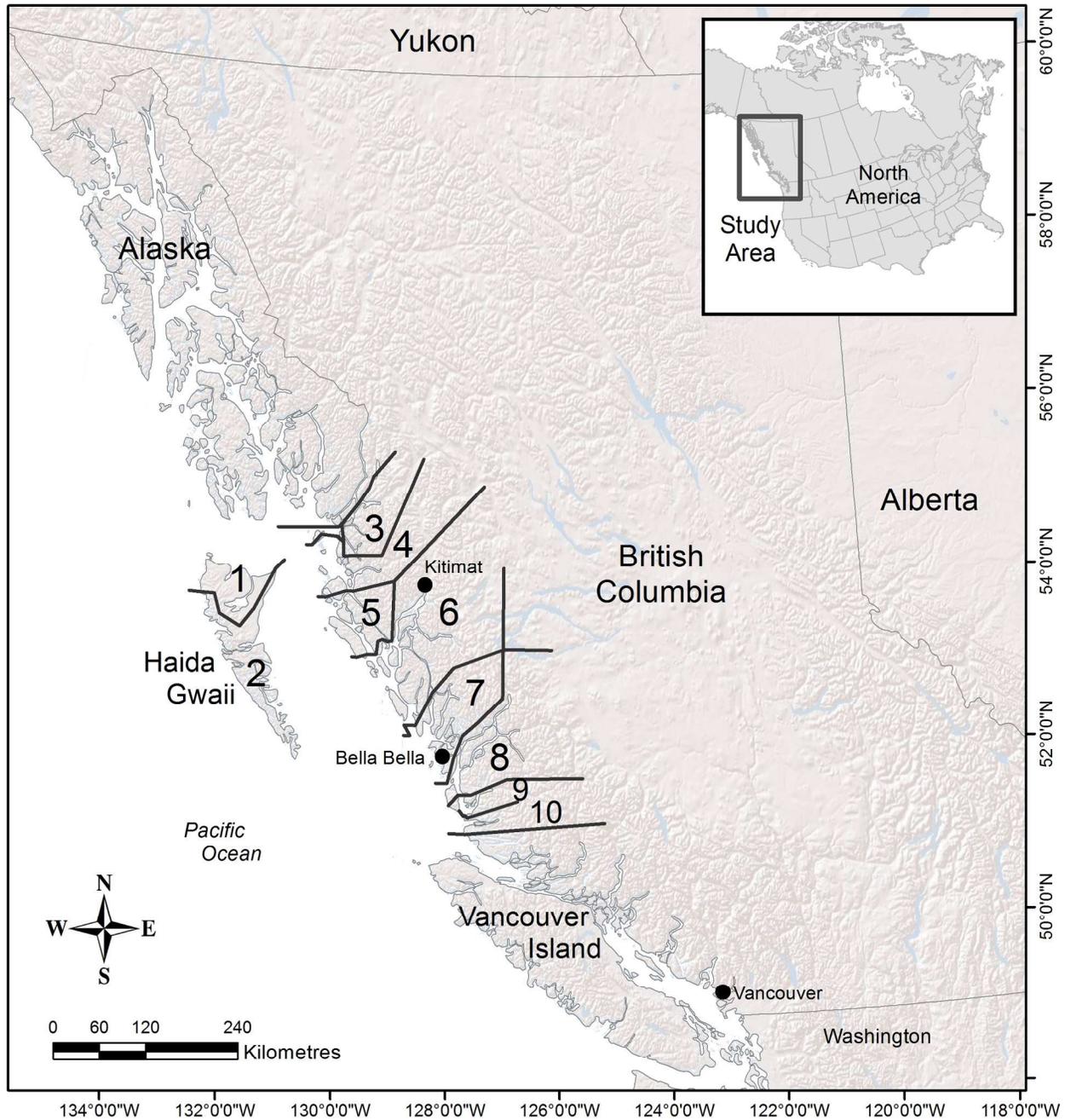
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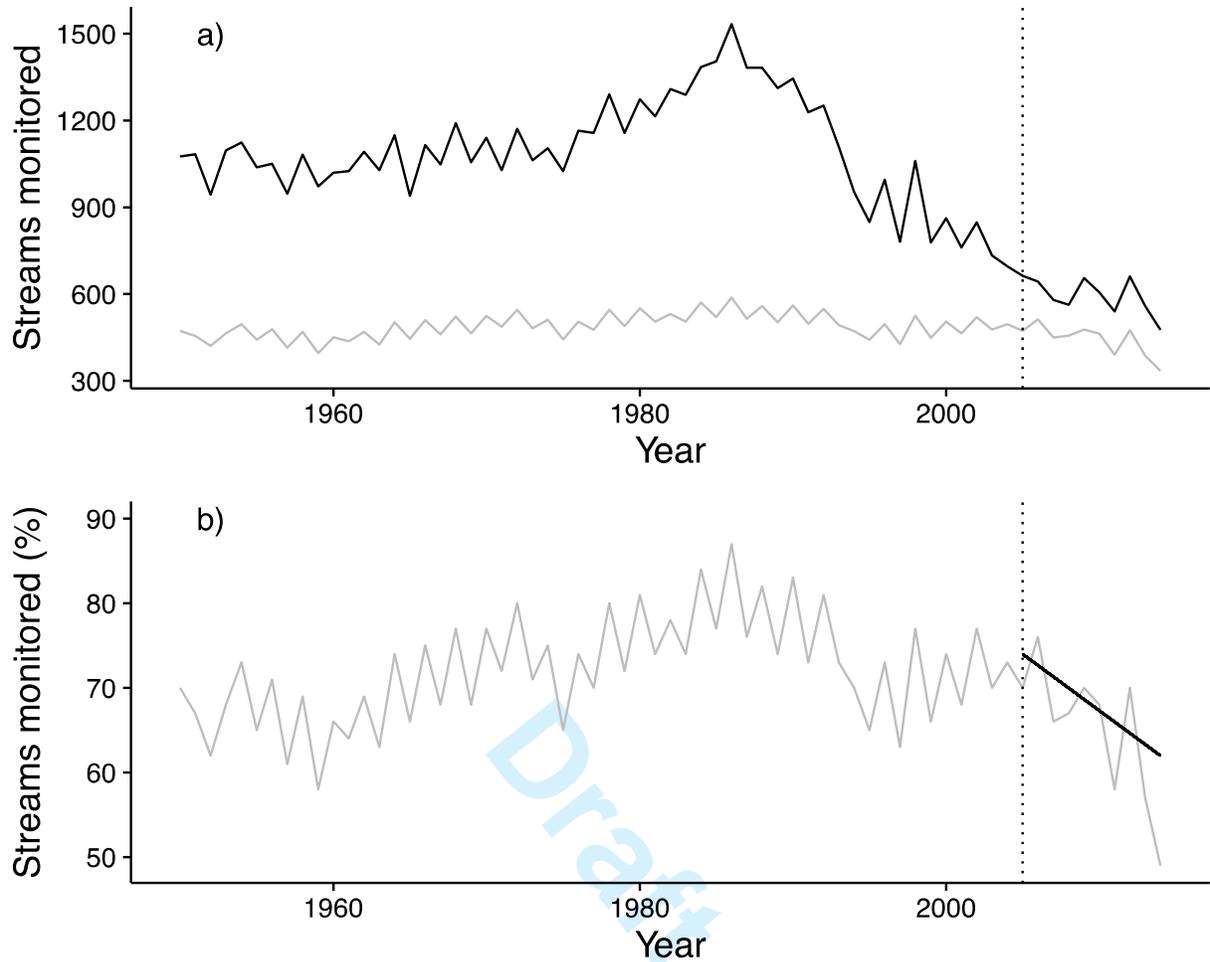
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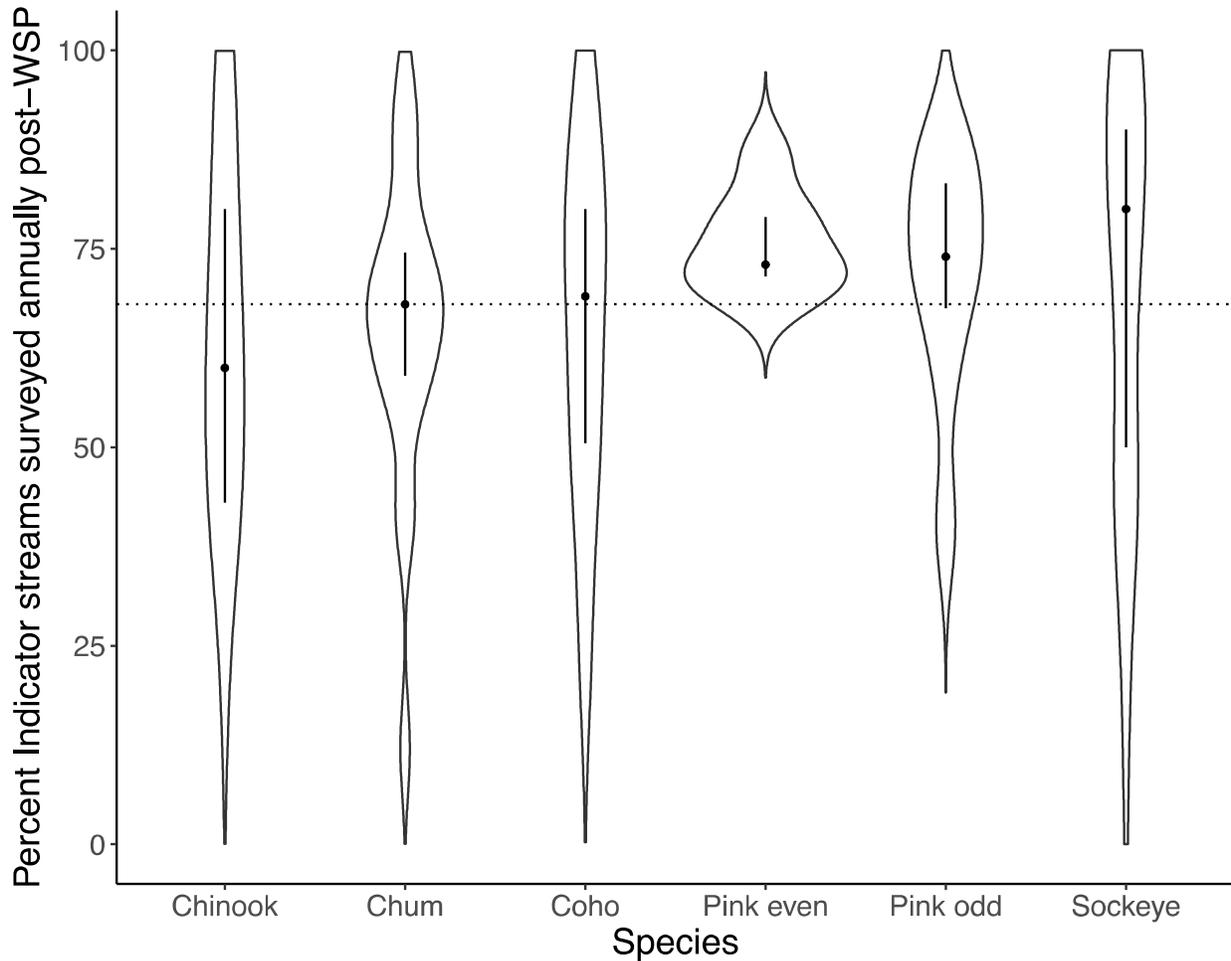
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**Figure 1.** Region encompassed by the Wild Salmon Policy, including British Columbia (BC) and Yukon Territory, and DFO Fisheries Management Areas 1-10 for Pacific salmon (*Oncorhynchus* spp.) returning to BC's north and central coasts where much of the data used in our analyses originate. Copyright 2014 Esri.



616  
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 618 **Figure 2.** a) Comparison of the number of spawning streams monitored between *all streams*  
 619 *combined* (black line) and *Indicator Streams* (grey line), and b) percentage of *Indicator Streams*  
 620 routinely (i.e.,  $\geq 50\%$  of the time) monitored, and the trend (black line) in monitoring since the  
 621 adoption of the Wild Salmon Policy in Management Areas 1-10 of British Columbia's north and  
 622 central coasts from 1950 to 2014. Vertical dashed lines demarcate the adoption of the Wild  
 623 Salmon Policy in 2005.

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632 **Figure 3.** Violin plots of the percentages of Indicator spawning streams surveyed annually  
 633 during 2005 to 2014, for each Conservation Unit (CU) with an assigned Indicator stream, for  
 634 each species along British Columbia's north and central coasts. Dashed line is the overall mean  
 635 monitoring effort across all CUs; black dots and lines are the medians and their 25<sup>th</sup>  
 636 percentiles.

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645 **Figure 4.** Violin plots of the relative change in average numbers of spawning adult salmon  
 646 (escapement) for each Conservation Unit within each species between the decade *Before* (1995-  
 647 2004) and *After* (2005-2014) the adoption of the Wild Salmon Policy. Dashed line demarcates  
 648 zero change in escapement; black dots and lines are the medians and their 25<sup>th</sup>  
 649 percentiles.

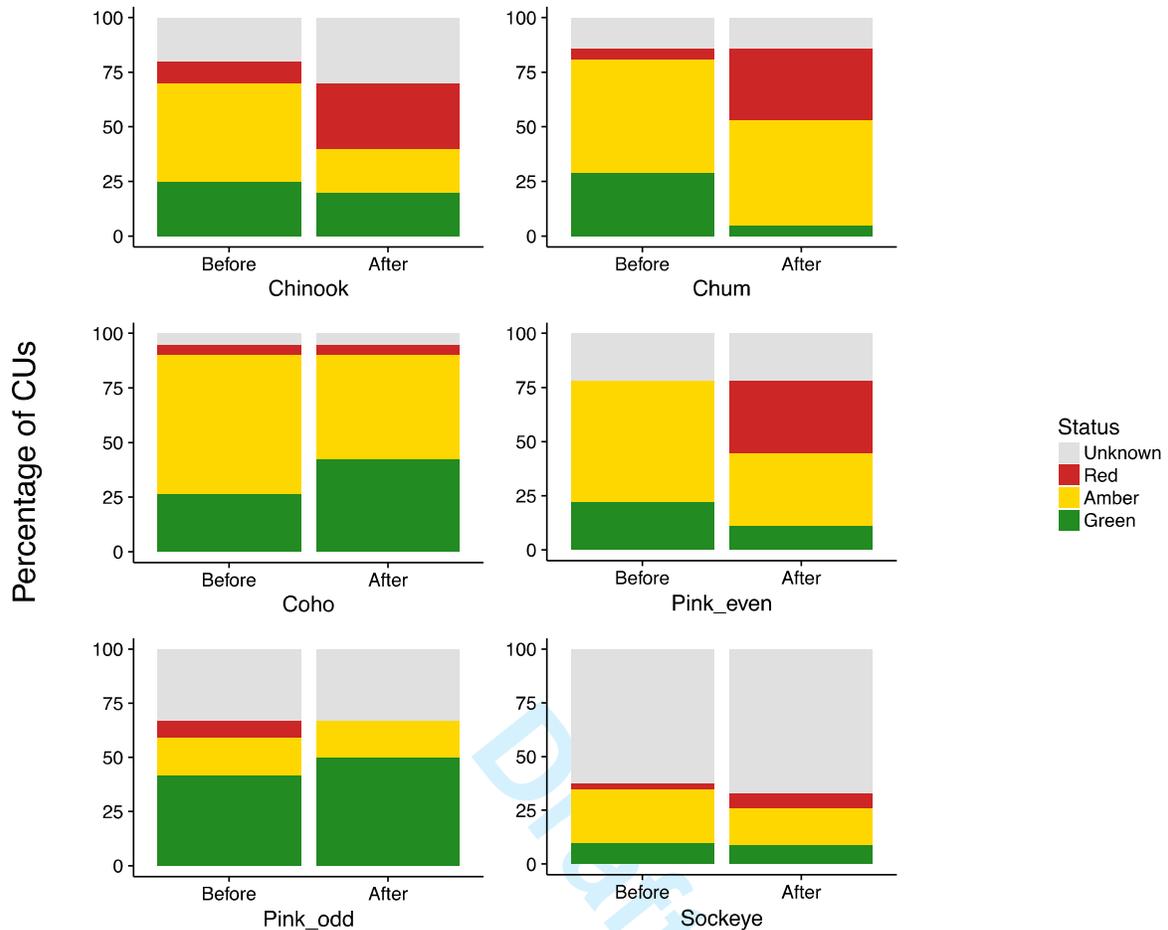
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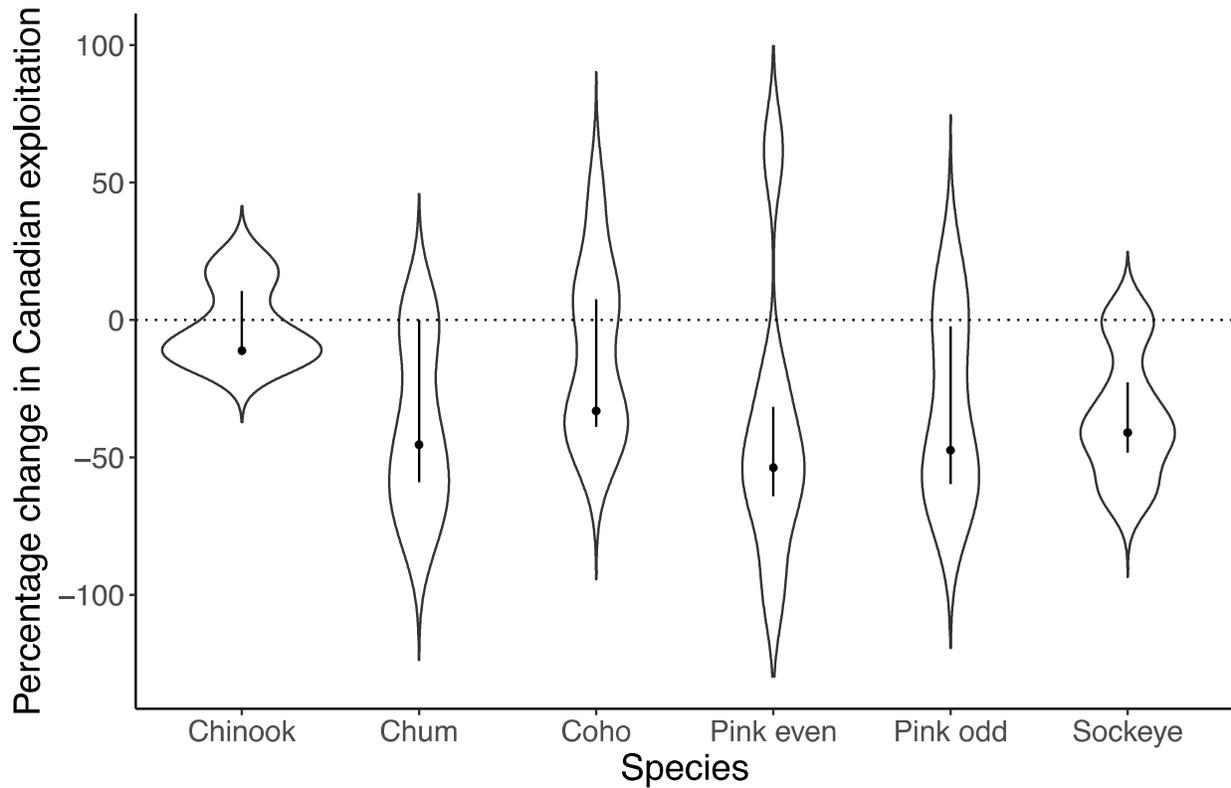


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657 **Figure 5.** Status assignments for all Conservation Units (CU) and species based on average  
658 escapement *Before* (1995-2004) and *After* (2005-2014) the adoption of the Wild Salmon Policy,  
659 derived from benchmarks assigned using the 25<sup>th</sup> and 75<sup>th</sup> percentile approach of historical  
660 escapement for each CU.

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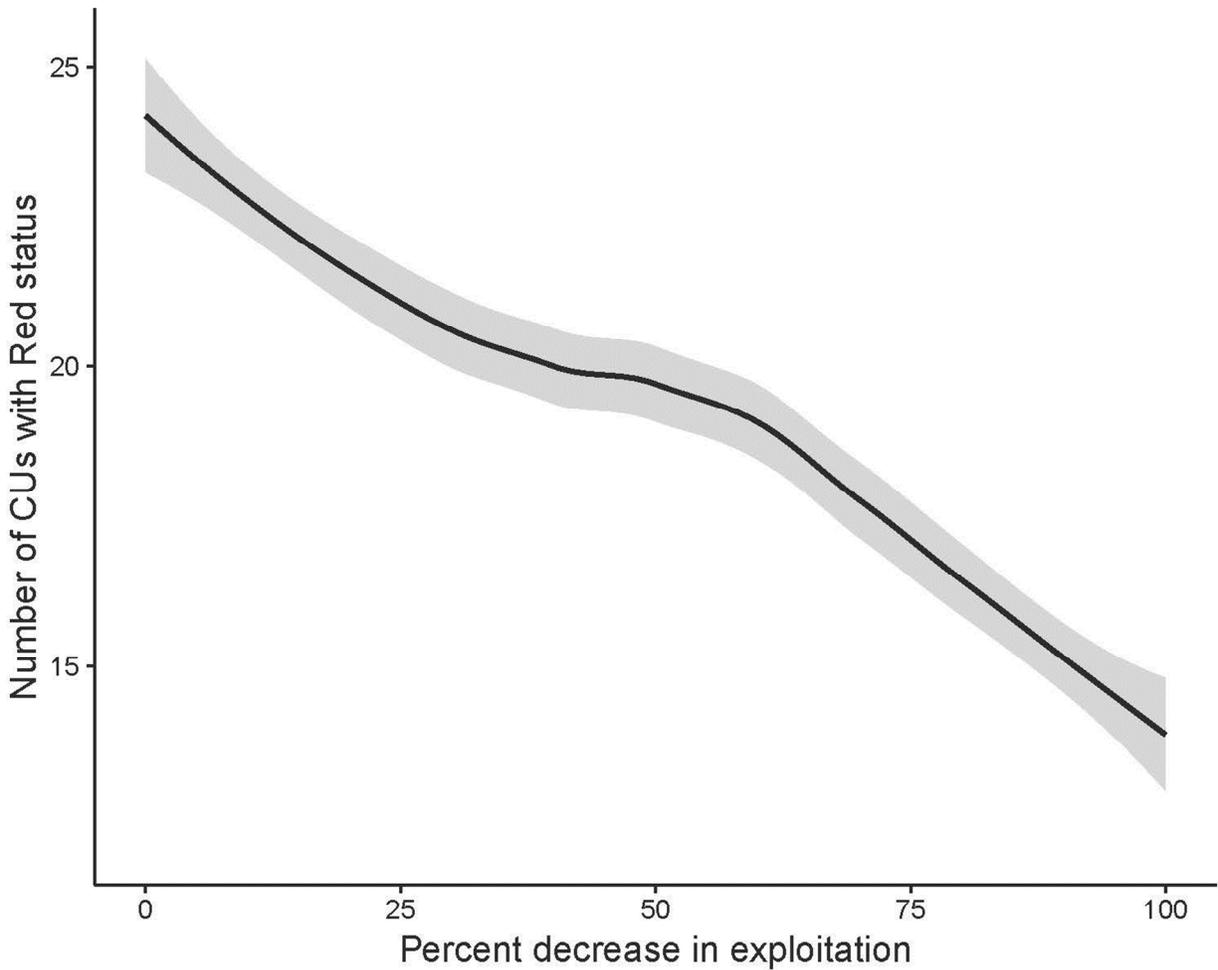


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676 **Figure 6.** Violin plots of the percentage change in Canadian exploitation rates on each  
677 Conservation Unit within each species between the decade *Before* (1995-2004) and *After* (2005-  
678 2014) the adoption of the Wild Salmon Policy. Dashed line demarcates zero change in  
679 exploitation rates; black dots and lines are the medians and their 25<sup>th</sup> and 75<sup>th</sup> percentiles.

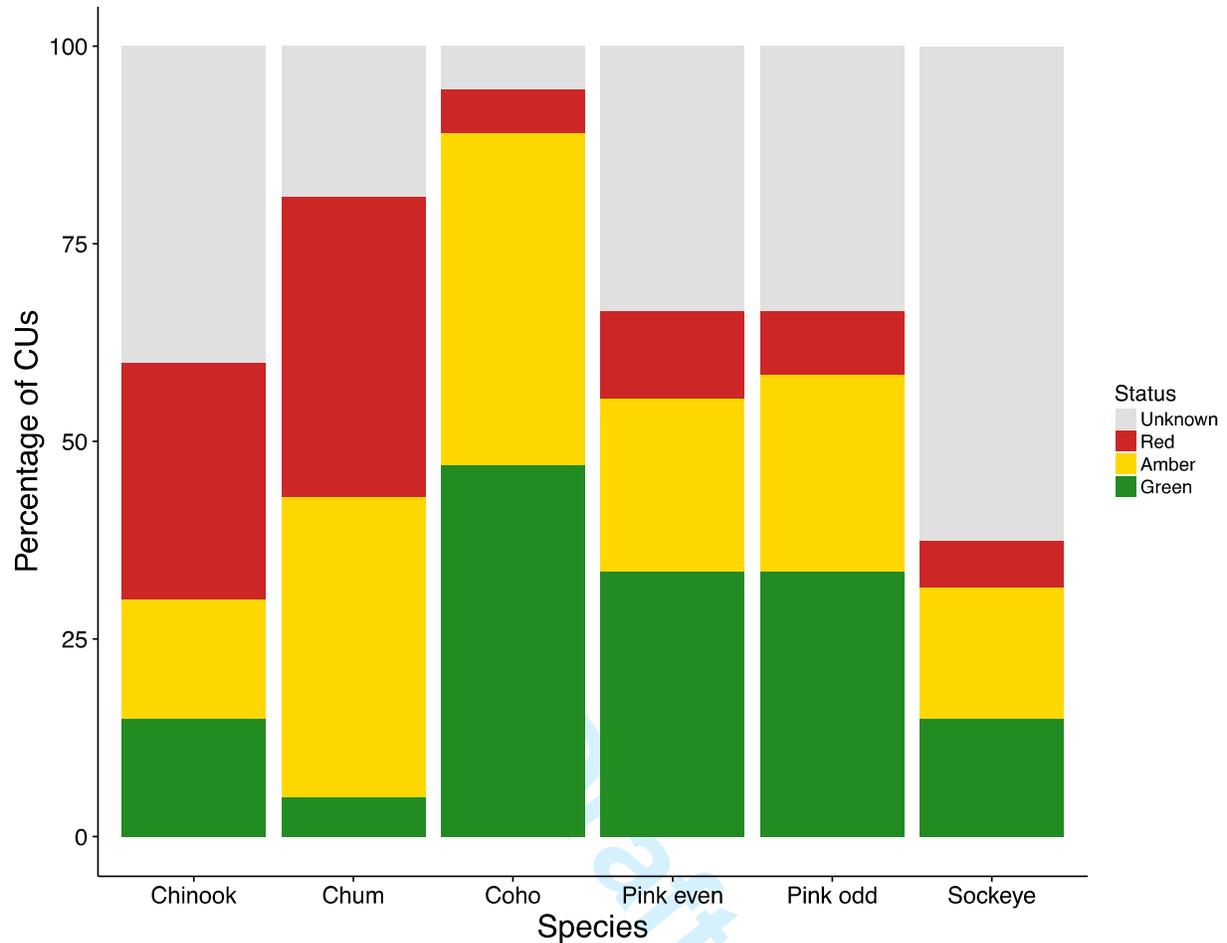
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695 **Figure 7.** Relationship between the number of Conservation Units (CU) assessed with Red-zone  
696 status and percent reduction in Canadian fisheries exploitation of each CU across all species on  
697 BC's north and central coasts in the decade after the adoption of the Wild Salmon Policy in  
698 2005. Black line is a fitted LOESS curve with standard error (grey shade).

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 707 **Figure 8.** Contemporary period status assessment of Conservation Units (CU) for each species  
 708 on BC's north and central coasts, based on the 25<sup>th</sup> and 75<sup>th</sup> percentile approach of spawner  
 709 abundance, and the following "generational" (i.e., over the most recent generation) years:  
 710 Chinook (2009-2014; 23 CUs); chum (2011-2014; 20 CUs); coho (2011-2014; 18 CUs); even-  
 711 year pink (2014; 8 CUs); odd-year pink (2013; 11 CUs); sockeye (2009-2014 depending on CU;  
 712 138 CUs).

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719 **Table 1.** Comparison of species-specific monitoring effort of spawning streams on BC's north  
 720 and central coasts between those recommended in the Core Stock Assessment Program (English  
 721 et al. 2006) and recent (2007-2014) efforts to determine the extent that a strategic approach has  
 722 been implemented since the adoption of the Wild Salmon Policy in 2005.

Species	Streams with annual surveys recommended*	Streams with annual surveys met*	Streams with periodic surveys recommended**	Streams with periodic surveys met**	Streams with annual and periodic surveys met (%)	Recommended streams with zero effort (%)
Chinook	33	8	28	6	23	36
Chum	175	35	42	20	25	11
Coho	104	17	101	25	20	51
Pink - even	157	60	N/A	N/A	38	11
Pink - odd	132	53	N/A	N/A	40	11
Sockeye - lake	27	12	6	2	42	18
Sockeye - river	6	2	3	0	22	56
Combined total	634	187	180	53	29	24

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\* Includes annual fence counts and mark-recapture programs

725 \*\* To be performed once every 2, 3, or 4 years, or 2 of 3 years

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729 **Table 2.** Synopsis of the change in spawning abundance, percentage of Conservation Units (CU)  
 730 that declined in spawning abundance, the number of CUs that declined to Red status, the  
 731 influence of Canadian fisheries on CUs assessed as Red, and the change in Canadian fisheries  
 732 exploitation on all salmon species returning to British Columbia's north and central coasts since  
 733 the adoption of the Wild Salmon Policy in 2005.  
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Species	Average spawning abundance 2005-2014	Change in spawning abundance	Percentage of CUs that declined in spawning abundance	Number of CUs that declined to Red status	Number of CUs that would have declined in status**	Number of CUs that would have improved in status***	Percentage change in Canadian fisheries exploitation
Chinook	98,000	-25,000 *	69	5	0	5	-2
Chum	1,135,000	-999,000 *	89	5	1	2	-11
Coho	1,137,000	237,000	35	1	0	3	1
Pink - even	6,485,000	-4,421,000 *	86	2	1	1	-10
Pink - odd	11,255,000	1,998,000	25	0	0	1	-
Sockeye	874,000	-104,000	53	8	1	5	-9
Combined total	20,984,000	-3,314,000	58	21	3	17	-7

735 \* Denotes statistical significance with Wilcoxon Signed-Rank Test for non-normal data.

736 \*\* Had fisheries exploitation not been reduced in the decade since the adoption of the WSP.

737 \*\*\* Had fisheries exploitation been further reduced in the decade since the adoption of the WSP.